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To cite this article: Salvatore Manfreda, Domenico Miglino, Khim Cathleen Saddi, Seifeddine Jomaa, Anette Eltner, Matthew Perks, Salvador Peña-Haro, Thom Bogaard, Tim H.M. van Emmerik, Stefano Mariani, Ian Maddock, Flavia Tauro, Salvatore Grimaldi, Yijian Zeng, Gil Gonçalves, Dariia Strelnikova, Martina Bussettini, Giulia Marchetti, Barbara Lastoria, Zhongbo Su & Michael Rode (21 Mar 2024): Advancing river monitoring using image-based techniques: Challenges and opportunities, Hydrological Sciences Journal, DOI: [10.1080/02626667.2024.2333846](https://doi.org/10.1080/02626667.2024.2333846)

To link to this article: <https://doi.org/10.1080/02626667.2024.2333846>



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Accepted author version posted online: 21 Mar 2024.



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Publisher: Taylor & Francis & IAHS

Journal: *Hydrological Sciences Journal*

DOI: 10.1080/02626667.2024.2333846

*Advancing river monitoring using image-based techniques:
Challenges and opportunities*

Salvatore Manfreda¹, Domenico Miglino¹, Khim Cathleen Saddi^{1,2,3}, Seifeddine Jomaa⁴, Anette Eltner⁵, Matthew Perks⁶, Salvador Peña-Haro⁷, Thom Bogaard⁸, Tim H.M. van Emmerik⁹, Stefano Mariani¹⁰, Ian Maddock¹¹, Flavia Tauro¹², Salvatore Grimaldi¹², Yijian Zeng¹³, Gil Gonçalves¹⁴, Dariia Strelnikova¹⁵, Martina Bussetini¹⁰, Giulia Marchetti¹⁰, Barbara Lastoria¹⁰, Zhongbo Su¹³, Michael Rode^{4,16}

- 1) Department of Civil, Architectural and Environmental Engineering, University of Naples Federico II, 80125 Naples, Italy
- 2) Istituto Scuola Superiore Pavia (IUSS Pavia), Pavia, Italy
- 3) Department of Civil Engineering and Architecture, Ateneo de Naga University, Naga, Philippines
- 4) Department of Aquatic Ecosystem Analysis and Management, Helmholtz Centre for Environmental Research—UFZ, 39114 Magdeburg, Germany
- 5) Institute of Photogrammetry and Remote Sensing, Technische Universität Dresden, 01069 Dresden, Germany
- 6) School of Geography Politics and Sociology, Newcastle University, Newcastle upon Tyne, United Kingdom
- 7) Photrack AG, Ankerstrasse 16a, 8004 Zurich, Switzerland
- 8) Water Resources Section, Faculty of Civil Engineering and Geosciences, Delft University of Technology, Delft, The Netherlands
- 9) Hydrology and Environmental Hydraulics Group, Wageningen University and Research, Wageningen, The Netherlands
- 10) Italian Institute for Environmental Protection and Research (ISPRA), Via Vitaliano Brancati 48, 00144 Roma, Italy
- 11) School of Science and the Environment, University of Worcester, Worcester, UK
- 12) Department for Innovation in Biological, Agro-food and Forest Systems, University of Tuscia, Viterbo, Italy
- 13) Faculty of Geo-Information Science and Earth Observation (ITC), University of Twente, 5 7514 AE Enschede, The Netherlands
- 14) INESC Coimbra, Department of Mathematics, University of Coimbra, 3030-290 Coimbra, Portugal
- 15) School of Geoinformation, Carinthia University of Applied Sciences, Villach 9524, Austria
- 16) Institute of Environmental Science and Geography, University of Potsdam, Potsdam-Golm, Germany.

Abstract

Enhanced and effective hydrological monitoring plays a crucial role in understanding water-related processes in a rapidly changing world. Within this context, image-based river monitoring has shown to significantly enhance data collection, improve analysis and accuracy, and support effective and timely decision-making. The integration of remote and proximal sensing technologies, with citizen science, and artificial intelligence may revolutionize monitoring practices. Therefore, it is crucial to quantify the quality of current research and ongoing initiatives to envision the potential trajectories for research activities within this specific field. The evolution of monitoring strategies is progressing in multiple directions that should converge to build critical mass around relevant challenges to meet the need for innovative solutions to overcome limitations of traditional approaches. The present study reviews showcases and good practices of enhanced hydrological monitoring in different applications, reflecting the strengths and limitations of new approaches.

Keywords: river monitoring; image-based techniques; remote sensing; water quality; artificial intelligence; citizen science

1 Introduction

Water resources management is facing critical challenges due to the combined effects of global warming, population growth, human pressures, and increased pollution. These factors collectively contribute to the global rise in hydrological extremes, including droughts and floods. Furthermore, they exacerbate the declining trend in water availability and degradation of water quality, which could ultimately result in chronic water scarcity affecting a substantial portion of the world's population. Already, around four billion people, approximately half of the global population, are affected by severe water scarcity (Mekonnen & Hoekstra, 2016), and future climate scenarios are expected to amplify this situation (Wheater & Gober, 2015; Lu et al., 2019; Trambly et al., 2020; Boretti & Rosa, 2019), which poses an escalating risk to human health and rights, ecosystems, cultural heritage, and the global economy (e.g., Cammalleri et al., 2020). In addition, water quality is being degraded more rapidly and diversely than ever with an increasing number of pollutants such as nutrients, pesticides, plastics, and emerging contaminants (Bhateria & Jain, 2016; Hannah et al., 2022). To address these challenges effectively, the field of water resources monitoring must evolve by considering the complex interconnections between the environment and human society (Montanari et al., 2013; Ross & Chang, 2020).

Although the commonly used and existing monitoring systems have laid the foundation for our knowledge, these have been designed under different hydrological conditions compared to today's needs and challenges.

These monitoring systems are laborious, expensive, and often provide discontinuous data in space and time (Sergeant & Nagorski, 2014). For example, current low-frequency water quality sampling methods fail to capture occasional pollution events or changes induced by urban floods, while high-frequency monitoring approach using in-situ water quality sensors is more likely to detect these dynamics in a detailed manner (e.g. Outram et al, 2014; Rode et al. 2016). However, these approaches are prone to instrument degradation (e.g., bio-fouling, calibration issues) and human error, if no adequate instrumental maintenance and technical staff training are ensured. Thus, it is necessary to adopt new observational strategies, benefiting from the increasing technological development, to deepen comprehension and gain further insights on river ecosystems and dynamics.

Advancements are expected to enhance the spatio-temporal resolution of observations in order to improve 'near real-time' water quality and quantity monitoring to move towards a more equitable, sustainable and efficient water management. In fact, water management practices face limitations concerning data availability and timely data sharing, especially in rapidly changing environments. Increasing in-depth knowledge of how climate change, as well as human pressures, impact the environment, both in time and space, cannot be achieved solely with traditionally used instruments. Furthermore, the absence of a data-sharing policy in certain regions, coupled with the lack of common standards and protocols for hydrological monitoring, metadata storage, and exchanges—especially for qualitative monitoring—still presents additional challenges to achieving effective and sustainable water management.

Recent advancements in Earth Observation (EO) technologies, environmental tracers (isotopes and biomarkers), new sensor technologies, and unmanned aerial systems (UAS), present promising opportunities to revolutionize environmental monitoring (Demarchi et al., 2017; Manfreda et al., 2018; Koparan et al., 2018; Wang & Yang, 2019; Koparan et al., 2020; Perks et al., 2020; Taramelli et al., 2020; Piegay et al., 2020; Strelnikova et al., 2023). Furthermore, the integration of ground-based measurements with remotely acquired data makes it possible to characterize environmental processes much more accurately than in the recent past. In recent years, there have been strong investments aimed at managing and maintaining EO missions, EO-derived services, and products, designing and launching new satellite missions or making operational new EO-based reliable tools (McCabe et al., 2017) and financing space research programmes. At the EU level, the European Copernicus Programme (<https://www.copernicus.eu/>) provides EO data and information services for different domains, and the EUMETSAT Satellite Application Facilities (SAFs) provides datasets and products for operational hydrological applications (referred to as H SAF; <https://hsaf.meteoam.it/>). Similar investments are also ongoing at national levels. For instance, Italy is pursuing investments in its space economy through the Mirror Copernicus Programme, focusing on national downstream services tailored to end user requirements. Part of this program, known as the IRIDE Program, is now being put into action as part of the Italian National Recovery and Resilience Plan. This initiative aims to enhance hydrological monitoring by launching a hybrid satellite constellation and providing EO services, with a particular focus on water management (Mariani & Bussettini, 2021). At global scale, a large effort was spent to accelerate EO uptake and impact by fully capitalising on the power of satellite EO in international development assistance operations such as the Global Development Assistance (GDA) program (<https://gda.esa.int/>). The GDA program is powered by the European

Space Agency (ESA) and implemented in partnership with the World Bank and the Asian Development Bank through complementary thematic areas, including water resources (<https://gda.esa.int/thematic-area/water-resources/>).

Like how smartphones revolutionized communication, the field of remote sensing has undergone a significant transformation with the emergence of miniaturized technology. In this context, CubeSats and UASs have boosted widespread adoption of these systems in academia, in operational institutional services, and in the commercial sector (Acharya et al., 2021; Eltner et al., 2022; Manfreda & Ben Dor, 2023). Therefore, there is a growing number of new users for such new technologies which still require harmonising effort in monitoring practices and accuracy assessment procedures (Tmusic et al., 2020).

The objective is to highlight the potential, limitations, and challenges of new technologies in hydrological monitoring by exploiting the capabilities of remote sensing, camera systems mounted on board of UAS or in fixed locations, image processing, and Artificial Intelligence (AI) algorithms. These methods and approaches may provide complementary and valuable information as well as processing capabilities to comprehensively monitor fluvial systems. This study is built upon the experiences of the authors who have been involved in European and national projects and are collaborating within the MOXXI Working Group of the International Association of Hydrological Sciences (IAHS – <https://iahs.info/Initiatives/Working-Groups/MOXXI/>).

2 Challenges in hydrological monitoring

Effective hydrological monitoring faces several challenges to ensure sustainable and equitable water resources management. This section will discuss three key challenges: data scarcity and institutional limitations, spatial and temporal variability, and increasing demands for water resources.

2.1 Data scarcity and institutional limitations

The primary challenge in hydrological monitoring is the data scarcity. Traditional monitoring systems often suffer from inadequate spatial coverage, limited temporal resolution, and insufficient availability of data. This scarcity of data hinders the accurate assessment of water resources and their quality, making it difficult to develop robust management strategies.

One of the most pressing concerns in data collection is the fragmentation of agencies and institutions responsible for overseeing distinct monitoring networks aimed at various objectives while tracking the same variables. This results in a heterogeneous and non-uniform distribution of monitoring stations, which often lack connections to a shared database or are installed at locations not suitable for specific objectives (Kirchner, 2006). Despite the overall increase in the number of sensors deployed over time, the availability of pertinent information has not shown significant improvement.

The development of monitoring systems over time have been significantly shaped by political decisions and mono-sectorial water management criteria that leads to fragmented networks. For instance, the Italian hydro-meteorological monitoring network, which transitioned from national to local control, has experienced relevant changes over time in the number of institutions involved, the number and distribution of monitoring stations (Braca et al., 2021).

In addition, it is not uncommon to observe the redundancy of investments in multiple monitoring networks carried out by various agencies (see the example of Basilicata Region in Fig. 1), with different purposes (e.g., hydrological monitoring, agrometeorological monitoring, civil protection, research). Even if these monitoring efforts have resulted in an increased number of monitoring stations, in-situ data fragmentation has increased over the course of time. These investments fail to enhance the quality and quantity of information provided because none of them are synchronized or optimized with each other. Due to the lack of coordination and data harmonisation, it is hard to have a comprehensive picture of this issue at a larger scale (e.g., at a national level or European level).

Therefore, the final total number of sensors that may be available over a certain area could comprise the combination of multiple networks established over the years, culminating in an exceedingly inefficient monitoring system. The heterogeneity of the monitoring networks also brings to light problems relating to the quality and officiality of the data.

The shift towards the digital age has unlocked the potential to construct extensive databases amalgamating measurements gathered from various origins. This could potentially overcome the existing constraints of current independently operational systems. However, substantial efforts are required for the reconstruction and harmonization of all available hydrological information, along with the digitization of historical data previously recorded on strip charts. In this context, certain automated techniques for data reconstruction have been formulated, which may help to reconstruct time-series reported on strip charts (e.g., Deidda et al., 2007; Jaklic et al., 2016).

Despite the significant number of rain gauges distributed worldwide, which sum to a total number ranging between 150,000 and 250,000, the heterogeneity of rainfall fields challenges their ability to comprehensively capture precipitation patterns (Groisman & Legates, 1995; Kidd et al., 2017). These gauges, assuming each is independent and represents a 5 km radius area, cover only about 1% of Earth's land surface (Becker et al., 2013). The number of river monitoring stations is smaller than rain gauges (e.g., in Italy, their number is approximately a third compared to approximately 3000 rain gauges available - Braca et al., 2021) and unevenly distributed across the globe. These stations are mainly concentrated in North America and Europe which represent about 50% of the global coverage, while Africa contains only 6% of the total (e.g., Herold & Mouton, 2011). In addition, water level stations only provide indirect measurements of discharge and require yearly surveys to reconstruct the corresponding updated flow rating curve (Mosley & McKerchar, 1993; Manfreda, 2018). This activity is time consuming and expensive and for this reason has been interrupted in several sites in recent years (Tauro et al., 2018a). Therefore, the real number of river monitoring stations useful for water assessment is even lower with respect to the number mentioned above.

Water quality monitoring is probably one of the most complex activities which frequently implies field sampling standards, complex laboratory protocols and techniques as well as routine data analysis. The Waterbase European Environment Agency (EEA) databases report on the status and quality of Europe's rivers, lakes, groundwater bodies and transitional, coastal, and marine waters, on the quantity of Europe's water resources, and on the emissions to surface waters from point and diffuse sources of pollution (Waterbase - Water Quality ICM, 2022). It shows that there are about 1,550 monitoring locations, distributed over 24

European countries, having three or more years of data with an average of at least four samples per year. This database represents just a subset of the EU Water Quality Monitoring Network, but the limited temporal resolution of most of the Water Quality observations does not allow to capture variability of natural and anthropic processes especially with respect to pollution events (Alilou et al., 2019). At a global scale, the Global Database of Freshwater Quality GEMStat (<https://gemstat.org/>) is one of the most comprehensive repositories of measured water quality data and gathered with voluntary submissions from different countries and organizations. The GEMStat database contains over 15 million entries from about 130,000 stations gathered from more than 80 countries (<https://gemstat.org/about/data-availability/>). Even if many gauging stations contain only a small fraction of available data, the GEMStat database represents an important open-access and valuable reference for in-situ water quality at global scale.

To address this limitation, the GlobeWQ project (<https://www.globewq.info/>) is a leading global initiative that proposes the integration of data from different sources, including in-situ, Earth observation (EO), and modeling results, to enhance water quality information and assessment globally. Additionally, the presence of macro- and micro-plastics in rivers poses one of the most critical challenges for ocean pollution. However, there are currently no standardized protocols and sustainable systems for monitoring it.

A recent study by Hurley et al. (2023) highlighted that the total number of monitored sites for macroplastics is limited to approximately 57 rivers around the world, which is notably irrelevant given the scale of the problem. Therefore, it is vital to enhance monitoring efforts in both space and time to better understand regional and global plastic fates.

2.2 Spatial and temporal variability

Another significant challenge is represented by the spatial and temporal variability of hydrological processes and water resources. Their patterns exhibit substantial variations and are influenced by factors such as climate, land use, soil characteristics, morphology, human activities, and interventions. Traditional monitoring systems, often based on pointwise measurements or sampling, struggle to adequately capture this variability. In fact, the hydrological variables are mutually influenced and controlled by the spatial variability of physical factors (e.g., Rodriguez-Iturbe et al., 2006; Metzger et al., 2017; Meijer et al., 2021). Therefore, water resources regimes can differ significantly between and within river basins due to the heterogeneity of geological settings, land cover, soil types, and human pressures. To account for this variability, monitoring networks must be designed to capture such heterogeneities. This requires the optimal distribution and densification of monitoring stations, and the use and integration of remote sensing data to gather spatially explicit information. This is also a clear objective introduced by the Water Framework Directive 2000/60/EC (WFD), although not always fully implemented, due to tangible limitations (e.g., insufficient funding, lack of skilled human resources).

Temporal variability also poses an additional challenge for water availability and quality monitoring. Infrequent sampling or sparse data collection fail to adequately capture water dynamics that can vary dramatically over different timescales ranging from hourly fluctuations to seasonal variations, and long-term trends. Therefore,

high-frequency monitoring, enabled by advanced sensor technologies and automated data collection systems, is crucial for accurately capturing these processes (e.g., Sergeant & Nagorski, 2014; Rode et al. 2016).

2.3 Increasing demands for water resources

Population growth, urbanization, and industrial development exert pressure on water availability and quality. This pressure is emphasized by the current and likely future impacts of climate change on water resources. Balancing the competing demands for water resources while ensuring their sustainable use and allocation requires monitoring networks which can be expanded and upgraded to provide comprehensive coverage and real-time data. However, traditional monitoring approaches often struggle to keep pace with the increasing demands for data.

Furthermore, as water scarcity becomes more and more pressing, efficient water management strategies are needed to optimize water allocation and minimize waste. Integrated monitoring systems that combine hydrological data with socio-economic information can facilitate informed decision-making and support sustainable water resources management.

Addressing the challenges of data scarcity and limitations, spatial and temporal variability, and increasing demands for water resources requires a concerted effort from the scientific community, policy- and decision-makers, and water resources managers. These challenges are clearly identified by IAHS Water Solutions Decade on “Science for Solutions: Hydrology Engaging Local People IN one Global world (HELPING)”. In this context, the theme 3 is promoting joint effort to integrate new technologies with existing ones (IAHS, 2023).

3 Advancing hydrological monitoring

Fast developing technologies such as remote sensing, UAS, advanced sensor networks, and wireless data networks offer opportunities to improve data availability and accessibility, and to collect data more efficiently and comprehensively. These technologies can also provide relatively high-resolution data over large spatial extents and properly capture temporal variations of hydrological processes. Integration of these technologies with data-driven approaches, such as artificial intelligence (AI), can enable more accurate and reliable hydrological monitoring.

This section will explore key areas such as: remote sensing and satellite-based technologies, sensor networks and citizen science.

3.1 Remote sensing

Satellite-based technologies offer the main advantage of a wide-area coverage, capturing information on various hydrological variables such as precipitation, evapotranspiration, soil moisture, and surface water dynamics (Rango, 1994; Chen & Wang, 2018, Pereira et al., 2019; Albertini et al., 2022). This data, depending on the satellite mission characteristics, can be obtained at reasonable regular time intervals and at different costs (some data are also available for free, such as the Copernicus mission), allowing for the assessment of temporal changes and the characterization of spatial patterns.

Numerous observation systems are tailored for hydrological research. Within NASA's 19 Earth science missions, 9 are notably pertinent to hydrology, including AQUA, ICESat-2, GPM, GRACE, PMM, SLAP, SMAP, SWOT, and VIIRS (NASA, 2023). The European Space Agency (ESA) has 4 missions relevant to hydrology: CryoSat-2, EUMETSAT satellites, Copernicus Sentinel-1 and Sentinel-2, and SMOS (ESA, 2023). ESA intends to launch the EarthCARE mission to enhance understanding of clouds and aerosols' role in solar radiation reflection. China has made substantial progress in Earth hydrology-related observation with the Fengyun and Haiyang satellite series, which focuses on meteorological observations and oceanographic monitoring and plans to also launch the Water Cycle Observation Mission (WCOM) (Shi et al., 2016).

In addition, various national and international initiatives aim to advance the intersection of Earth observation and hydrological science. These include the International Precipitation Working Group (IPWG), NASA Energy and Water Cycle Study (NEWS), European Union WATER and Global CHange (WATCH), and the Global Energy and Water Exchanges (GEWEX, 2018) initiative.

The combined efforts of Earth observation missions and initiatives are propelling hydrology into the era of "Big Data" (Peters-Lidard et al., 2017). Big data techniques can handle vast amounts of data and extract meaningful insights and interpret complex hydrological datasets, leading to improved understanding and predictive capabilities.

UAS, alongside satellites, are also valuable tools for hydrological monitoring (Manfreda & Ben Dor, 2023). These systems use advanced sensors to collect high-resolution data on a local scale, allowing for precise observations. UAS are versatile and agile, capable of capturing RGB and multi- or hyper-spectral data, thermal imagery, and LiDAR (Light Detection and Ranging) data. They excel in covering from local to large areas, reaching inaccessible regions and improving the spatial and temporal resolution of hydrological observations. Currently, one limitation for the use of UAS is still represented by the maximum extent of individual surveys. Swarm drone operations could overcome these limitations using multiple unmanned aerial flying platforms carrying sensors integrated as a single networked system.

In this context, an intermediate level between satellites and UAS is represented by 'CubeSat' and High-Altitude Pseudo Satellites (HAPS), which offer the possibility of conducting surveys over larger surfaces with high resolution. CubeSats are a standard size and form factor nanosatellites which are increasingly utilised by space agencies (McCabe et al., 2017). HAPS, which typically fly at 15,000-30,000 m above ground level for several months at a time, are also currently in development and testing. Both systems have the potential to fill the gap between satellites and UAS for earth observation and hydrological monitoring, given their endurance and spatial resolution (Poghosyan & Golkar, 2017; Gonzalo et al., 2018; Baraniello et al., 2021).

3.2 Internet of Things (IoT) and sensor networks

The Internet of Things (IoT) represents a novel technological paradigm conceptualized as a worldwide network of machines and devices with the ability to engage in mutual interactions (Lee & Lee, 2015). It allows the integration of sensors interconnected through a variety of access networks, facilitated by cutting-edge technologies like embedded sensing and actuation, radio frequency identification (RFID), wireless networks, and semantic and real-time web services. The IoT's low-power wide area network (LPWAN) capabilities enable

the utilization of battery-powered sensors. Among these, Long Range Wide Area Network (LoRaWAN) stands out as it employs open-source technology and operates on unlicensed frequency bands, offering significantly greater range compared to WiFi or Bluetooth connections. LoRaWAN is particularly advantageous for applications in remote regions where cellular networks experience limited coverage. Given the extensive multitude of devices that can be connected, the number of local measurements may enormously increase offering the possibility to further explore complex dynamics of hydrological forcings (Perumal et al., 2015; McCabe et al., 2017; Balsamo et al., 2018; Tauro et al., 2018a; Tosi et al., 2020; Livoroi et al., 2021).

Sensory networks, in conjunction with the Internet of Things (IoT), hold promising potential for adaptive monitoring strategies through dynamic adjustments to the spatial distribution and density of sensors based on evolving hydrological conditions. Such flexibility enables targeted data collection in response to specific events or areas of interest, incorporating event-based triggers for sensor readings. Consequently, this approach optimizes the allocation of monitoring resources (Zanella et al., 2023; Marino et al., 2023).

3.3 Citizen science and crowd-sourced data

Citizen science represents an innovative tool with a measurable influence on environmental stewardship and public policy advocacy. When integrated with crowdsourcing, involving the gathering of extensive data from a diverse array of participants, citizen engagement can lead to an immediate, real-time response to pressing issues, such as pollution and natural disasters. Engaging the public in data collection and classification does not only increase data coverage, but also promotes public awareness and participation in water resources management (e.g., Nardi et al., 2022, Sermet et al., 2020).

Recently, several initiatives have stimulated the participation of volunteers in data collection through methods such as mobile applications, community-based monitoring programs, or distributed sensor networks, as is the case, for example, with the CrowdWater project (Strobl et al., 2019). Participants can measure water levels, discharge, report on water quality observations, and share hydrological data collected from their personal monitoring stations. This collaborative effort significantly boosts data availability and offers valuable insights into the specific hydrological conditions within local areas, while also providing an avenue for active participation of key stakeholders in the community to foster technology localization and sustainability. The use of crowd-sourced data also holds the potential to complement conventional monitoring networks by capturing detailed spatial and temporal data (e.g. Etter et al., 2020; Mapiam et al., 2022).

In this context, citizen-acquired data may not always be comparable to professionally obtained data. To maintain high data quality standards, initiatives in citizen science consistently incorporate training and implement measures for quality control. This enables participants to contribute effectively, following simple and reliable procedures. Integration of these data with professional monitoring data, remote sensing observations, or model simulations can notably enhance the accuracy and resolution of hydrological analyses. Several private sensor networks have been growing significantly in the last few years offering many measurements that may be easily filtered and validated offering a dense network of observations (de Vos et al., 2019). Popular online platforms such as Netatmo (<https://weathermap.netatmo.com>) and Weather Underground (<https://www.wunderground.com/>) collect and visualize measurements from public and even

personal weather stations (PWSs) every ~5 to 10 min with a total number of sensors that exceeds one order of magnitude the number of stations of the national hydrological services (Graf et al., 2021; Coney et al., 2022). These private network of rainfall stations connected to a web service has been clearly assessed in a recent work by Graf et al. (2021). In fact, these opportunistic networks can lead, after filtering and sensor calibration (Krüger et al., 2023), to rainfall maps of higher accuracy and increased spatial variability, especially on smaller spatial and temporal scales.

The successful use of citizen science's potential for hydrological monitoring relies on a crucial collaboration between scientists, water managers, and the public. This engagement with the public also proves to be an effective approach for fostering a deeper understanding on water level measurements, discharge estimates, water conservation and shoreline changes (Harley et al., 2019; Seibert et al., 2019; Wang et al., 2022).

Several citizen science initiatives and crowd-sourced data collection platforms can contribute to data integration efforts (for instance, the largest people-powered research platform Zooniverse - <https://www.zooniverse.org/>). Engaging the public in data collection and monitoring processes can increase data coverage and improve community involvement in water management decisions. This approach has been used for instance for: for flood risk management in Argentina, France, and New Zealand (e.g., Le Coz et al., 2016) or coastal litter monitoring (e.g., van Lieshout et al., 2020).

4 The potential of image-based techniques for river monitoring

The proliferation of modern optical sensors, present in satellites, on UAS, and in smartphones as well as attached to low-cost single board computers or micro-controllers, has sparked a compelling interest in utilizing imagery to greatly broaden the scope of possible hydrological observations (e.g., Manfreda et al., 2018; Strelnikova et al., 2023). Digital cameras have been successfully used in areas such as surveillance, facial recognition, object detection and tracking, inventory monitoring, and management. Additionally, some existing algorithms developed for general purposes have found intriguing applications in environmental monitoring, including the creation of 3D models (James et al., 2019), assessment of highway vehicle flux (Hsu et al., 2003), monitoring air pollution (Zhang et al., 2016), measuring rainfall intensity (Allamano et al., 2015; Kavian et al., 2018; Jiang et al., 2019), and mapping river velocity fields (Johnson & Cowen, 2017; Lewis & Rhoads, 2018), among others.

Numerous researchers are currently delving into the field of river monitoring using image processing techniques. This includes a range of possibilities like traditional image processing methods from computer vision, but also new techniques utilizing Artificial Intelligence (AI). 3D reconstruction through Structure-from-Motion and Multi-View Stereo (SfM-MVS) or photogrammetry is one of the most common applications.

It is becoming increasingly common for ordinary hydrological stations to be associated with nearby installed optical cameras. New camera systems have the potential to capture a range of complementary information (e.g., water level, velocity, and water quality parameters) useful for interpreting natural phenomena. They may also be exploited to measure features with the support of an evolving range of software and tools (e.g., Segment Anything Model - <https://segment-anything.com/>). In Figure 2, we provide three examples of images taken on different rivers highlighting various phenomena such as a drought, plastic pollution, and a flood with

associated intense wood transport.

The following sections provide an in-depth exploration of various aspects that can be investigated using image-based techniques, ranging from 3D reconstruction for analysing river morphology to monitoring water quality in rivers.

4.1 River morphology

Structure from Motion (SfM) with Multi-View Stereo (MVS) algorithms have produced a revolution in the field of high-resolution topographic reconstruction (Westoby et al. 2012). This integrated approach utilizes matching features in multiple overlapping images without requiring 3D location of ground control points, which is common in traditional photogrammetric methods. These methods may help estimating the area of the water surface in orthophotos (Niedzielski et al., 2016) and assessing the river morphology throughout time (Carrivick and Smith, 2019) allowing for frequent cross-section updates, which becomes especially important before and immediately after flood events (Bertalan et al., 2023). These surveys can be repeated several times over the year to define the cross-section geometry. To improve the change detection accuracies, multi-temporal image matching techniques can be used (Feurer et al., 2018) and when combined with error propagation methods, such as M3C2-PM (James et al., 2017), enable the identification also of small-scale events (Blanch et al., 2021). The magnitude of morphological events that can be identified with change-detection techniques depends on the resolution of the image and on the characteristics of the sensors. River morphological changes detection from satellite data (e.g. radar data) are not always easy to meet, showing limitations, potentials, and challenges in this field of study, still to be investigated (Marchetti et al., 2023; Rossi et al., 2023). The riverbed below the water surface can be reconstructed, using active or passive mapping approaches, and considering radiometric and geometric principles (Mandlbürger, 2022). Geometric tools consider refraction impacts (Maas, 2015; Dietrich, 2017) and radiometric tools utilize the attenuation of light when travelling into deeper water (Flener et al., 2013; Legleiter, 2021; Mandlbürger et al., 2021). Thus, images can be used to describe the observed river cross-section above and below the water surface (Eltner et al., 2021a), including the change of grain size distributions of the exposed river sediment bars (Lang et al., 2021; Marchetti et al., 2022) or growth of submerged vegetation (e.g., Rowan and Kalacska, 2021) which controls hydraulic roughness and conveyance in many lowland rivers or the change in geomorphic unit assemblages, which compose the riverine habitats (Carbonneau & Bizzi, 2023; Latella et al., 2024; Crivellaro et al., 2024).

4.2 Image velocimetry

Image sequences can be used to trace flow velocities and paths tracking and measuring the displacement of visible structures on the water surface. Generally, natural patterns on free surfaces, such as wave crests, vortices, bubbles, foams, or natural floating material (debris, vegetation) provide seeding for image processing. These conditions are especially prevalent during floods due to the presence of surface ripples caused by near-surface turbulence and pressure fluctuations. In case of missing tracers, they can also be added manually if the site conditions allow for it.

In recent years, several image velocimetry methods have been developed, which include the classical

correlation-based algorithms such as Large Scale Particle Tracking Velocimetry (LSPTV - Brevis et al., 2011; Tauro et al., 2017; Eltner et al., 2021a) and Large Scale Particle Image Velocimetry (LSPIV - Fujita et al., 1998; Muste et al., 2008; Sabrina et al., 2021). Other approaches measure patterns of image intensity in 1D, i.e., Space-Time Image Velocimetry (STIV - Fujita et al., 2007) or use the well-known computer vision technique of optical flow, e.g. implemented in the tools Optical Tracking Velocimetry (OTV - Tauro et al., 2018b) and Kanade–Lucas Tomasi Image Velocimetry (KLT-IV - Perks, 2020).

These methodologies have demonstrated their successful implementation in continuous monitoring systems (e.g., Peña-Haro et al., 2021) and they have been effectively applied in remote or hard-to-reach areas using footage captured from UAS or on synthetic flow videos (e.g., Perks et al., 2016; Tauro et al., 2016; Pearce et al., 2020; Ljubicic et al., 2024; Bodart et al., 2022). The introduction of deep learning-based algorithms, as highlighted by Ansari et al. (2023), has significantly reduced the need for extensive parameter configuration. It is important to note that the accuracy of these methods in reconstructing surface velocities is influenced by various factors, including camera setup, seeding density, lighting conditions and other environmental variables. Despite the associated errors being typically quite small, often lower than 10% (Manfreda et al., 2019; Eltner et al., 2020), data collection can still be improved by taking into consideration relevant data optimization procedures. Since there is an emphasis on the need for higher seeding density, solutions to optimize frame selection using metric-based (Pizarro et al., 2020) and Artificial Neural Network (ANN) (Alongi et al., 2023) approaches may be beneficial to improve the quality of surface velocity reconstructions.

Besides the usage of RGB imagery, thermal data is considered for the tracking tasks, which can become important in the case of low density or absence of floating features at the water surface (Lin et al., 2019) or in night-time conditions (Fujita, 2017). Thermal cameras can capture the inherent fluxes of river surface temperature and therefore trace the evolving vortices (Kinzel and Legleiter, 2019; Eltner et al., 2021c). Furthermore, these cameras have the potential to serve as valuable tools for monitoring the inflow of water with distinct thermal properties into the main river channel.

4.3 River discharge

The computation of river discharge primarily relies on integrating water velocity profiles with cross-sectional area. In this context, a crucial factor is the water level, which can be determined in various ways. One option is the imaging of a stage board for a straight-forward water level retrieval (Young et al., 2015; Leduc et al., 2018), or using synchronised cameras to 3D reconstruct the water surface (Ferreira et al., 2017). In ephemeral and intermittent streams, water level was monitored by applying automatic image thresholding to pictures of a reference thin pole installed in the stream (Noto et al., 2022; Tauro et al., 2022). Another approach is the masking of the water area in the image to identify the boundary between water and other elements and then transforming this waterline to water level data. The process involves two steps: image segmentation, which classifies pixels as water or non-water using image sequences or CNNs (Stumpf et al. 2016; Elias et al., 2019; Vandaele et al., 2021); and water level retrieval, where the waterline contour is converted into water level data, often through the intersection with a 3D model (Eltner et al., 2021b).

Once water levels and velocities are determined, discharge can be determined with various approaches. One

straightforward method involves deriving depth-averaged velocity with correction parameters, subsequently applying the velocity-area method (Hauet et al., 2008; Detert et al., 2017). Alternatively, the entropy model reconstructs flow velocity profiles (Bahmanpouri et al., 2022), utilizing known conditions like surface velocity to accurately estimate average water flow speed (Moramarco et al., 2013). A more recent approach uses river wave patterns to directly estimate discharge, without the need for cross-section measurements or velocity data, solely relying on physical principles (Dolcetti et al., 2022). The last methodology may be useful during high flood flows, but it still provides errors larger than 30%.

4.4 Water quality

Numerous water quality (WQ) parameters serve as key pollution indicators. Gholizadeh et al. (2016) identifies a list of most commonly measured qualitative parameters of water, which include: chlorophyll-a (CHL-a), Secchi Disk Depth (SDD), Temperature (T), Colored Dissolved Organic Matters (CDOM), Total Organic Carbon (TOC), Dissolved Organic Carbon (DOC), Total Suspended Matters (TSM), Turbidity (TUR), Sea Surface Salinity (SSS), Total Phosphorus (TP), Ortho-Phosphate (PO₄), Chemical Oxygen Demand (COD), Biochemical Oxygen Demand (BOD), Electrical Conductivity (EC), and Ammonia Nitrogen (NH₃-N). Some, like CHL and CDOM, have optical properties detectable with RGB cameras (Godijn & White, 2006), while non-optical parameters like Total Phosphorus (TP) can be remotely sensed by leveraging their relationship with optically active parameters, such as CHL (Niu et al., 2021).

Furthermore, RGB cameras are effective in detecting floating materials, plumes, foam, or oil spills, with spatial image resolution matching object dimensions being the only limitation. Feature detection and labeling algorithms provide valuable insights into material density and distribution, aiding in prompt pollution event detection.

A recent review by Blanco Ramirez et al. (2023) underscores the potential of citizen science in diverse hydrological applications, especially pollution detection and water quality modelling. The review highlights the need for guidelines and protocols to ensure data meets water quality standards and is comparable across projects.

Additionally, integrating discharge and flow velocity measurements enhances pollutant concentration and propagation velocity evaluation.

4.4.1 River turbidity

Turbidity detection is an important indicator of WQ and assumes great significance for environmental protection and aquatic ecosystems. It is used as a relative indicator for other physical properties such as suspended sediment concentration (SSC) and total suspended solids (TSS) as well as other compounds present in water such as chlorophyll, organic matter, microorganisms, algae, or chemicals, etc. Since turbidity is generally closely related to these compounds, it can often be used for quantitative estimation. Turbidity often varies with seasons and rainfall events that can cloud the water. The growth of algae and other organisms in the summer can also cause an increase in turbidity.

There are several applications of river turbidity monitoring from satellites which provide a clear overview

about opportunities offered by different spectral indices (e. g., Lacaux et al., 2007; Wang and Shi, 2007; Fraser, 1998; Constantin et al., 2016). Among others, Garg et al. (2020) investigated the change in spectral reflectance of water across the visible to NIR range along the Ganga River. The temporal analysis indicates a reduction in reflectance across the visible to NIR range, likely due to decreased water turbidity. While the blue and green bands struggle to map turbidity variations because of bottom interference, the red and NIR bands prove more sensitive for turbidity estimation, particularly in optically deep water. Ehmann et al. (2019) also confirmed these findings, emphasizing the red band's sensitivity in depicting turbidity gradients within UAS imagery.

4.4.2 Macroplastics and litter

The global concern about plastic pollution in rivers is increasing due to the escalating issue of plastics entering waterways through direct disposal, wind dispersal, runoff, and sewage discharge. Macroplastics, which are large particles of plastic debris (>2.5 cm), account for about 70 to 80 % of total debris that mainly enters the oceans via rivers, posing serious threats to the environment and human health (Haseler et al, 2018; GESAMP, 2019; van Emmerik & Schwarz, 2020). This concerning issue is further compounded by an annual surge of 14 million tons of plastic leakage at global scale. Notably, the yearly discharge of mismanaged plastic waste (MPW) from land into the ocean was estimated to vary between 0.41 up to 12.7 million metric tons (Jambeck et al., 2015; Lebreton et al., 2017; Schmidt et al., 2017). The range of variability is relatively large given the high uncertainty and low availability of reliable observations (Roebroek et al., 2022; González-Fernández et al., 2023). It is also remarkable that the distribution of emissions is extremely uneven with few countries such as Philippines and India that contribute up to almost 50% of the pollution (Meijer et al., 2021). Recent work however suggests that a large part of plastic accumulates in and around rivers for years to decades before reaching the ocean (van Emmerik et al., 2022). Currently, the plastic mobilization, transport and accumulation processes remain largely unresolved.

Flood events seem to play a key role, and globally lead to a 30-fold increase in plastic mobilization (Roebroek et al., 2021). Plastic transport can vary one to two orders of magnitude at daily, monthly, and yearly timescales. However, current methods only allow for sporadic observations with limited temporal and spatial coverage (van Calcar & van Emmerik, 2019; González-Fernández et al., 2021).

Despite the low reflectance of plastic, RGB and multispectral imagery combined with machine learning tools (e.g., Gnann et al., 2022) or in some cases with simpler tools, such as object detection (van Lieshout et al., 2020), Spectral Angle Mapper (Gonçalves and Andriolo, 2022), Naïve Bayesian classification (Biermann et al., 2020), image masking (Schreyers et al., 2021), and other RGB markers such as microorganisms (Corbari et al., 2020), are valuable tools for effective macroplastic monitoring. To overcome the large uncertainties in plastic transport, retention and emissions into the ocean, increased monitoring efforts are crucial.

5 Experimental activities and projects

The large number of ongoing international initiatives highlights the worldwide effort aimed at developing new tools and methodologies for hydrological monitoring. In Table 1, we have listed the most promising recent research projects that exploit innovative technological solutions, including image processing, UAS, and citizen

science.

Within this context, we are carrying out several experimental initiatives within international projects that are useful to underline the potential of alternative techniques with practical examples. Among others, we will mention: 1) ASI-ISPRA Habitat Mapping project; 2) the PRIMA-funded project named “OurMED: Sustainable water storage and distribution in the Mediterranean” and the 3) “RiverWatch: a citizen-science approach to river pollution monitoring” funded by Italian Ministry of University and Research PRIN. The first project aims to use remote sensing techniques for mapping terrestrial, aquatic, and transitional habitats, including characterizing fluvial hydromorphology. The second project aims to explore new water saving strategies over several demo sites distributed in different countries in the Mediterranean. Within each demo site, monitoring camera systems will be adopted to measure the water flow and turbidity. The third project will be focused on the monitoring of the Sarno River which is the most polluted river in Europe (Lofrano et al., 2015; Baldantoni et al., 2018). These examples trace the pace for a new trajectory in hydrological monitoring. These projects offer a unique opportunity to test new water monitoring ideas and tools.

More details about both case studies are given in the following with the aim to identify the expected impacts associated to the ongoing activities.

5.1 Hydromorphological characterization through satellite and UAS data

The ASI-ISPRA Habitat Mapping project (Table 1) established the “Italian Research and development Initiative for Spaceborne river monitoring” (IRIS) to develop a prototype of tools and algorithms able to map river forms and processes along hundreds of kilometres of river lengths. The combined use of Copernicus Sentinel-1 and Sentinel-2 data and UAS acquisitions (the latter used as “ground truth”, that is as the real feature collected at the imaged location) in a satellite-based classification algorithm has proven effective to identify, along the river corridor, the spatial units with similar textural and spectral characteristics that constitute the key river geomorphic macro-units: “water”, “vegetation” and “exposed sediment” (Fig. 3, Carbonneau et al., 2020; Mariani & Bussetini, 2021). These key geomorphic macro-units represent the external envelope of river geomorphic units of the same type (Belletti et al., 2017), whose temporal evolution is necessary to identify river channel dynamics, better understand river processes and their future trajectories.

The IRIS prototype was developed and tested on five Italian rivers characterized by different channel patterns, morphological types, and flow regimes, namely the Po and Sesia Rivers (in NW Italy), the Tagliamento River (in NE Italy), the Paglia River (in Central Italy), and the Bonamico Torrent (in Southern Italy). The operational implementation of these algorithms and tools is underway within the IRIDE Program (Table 1). The potential of this tool is to have continuous river monitoring and support the local authorities in the assessment of rivers morphological status, needed for the WFD implementation.

The satellite-based algorithm developed in IRIS is typically applied to a mask of the river corridor, divided into specific river stretches of similar morphological types, following the IDRAIM methodological framework for stream hydromorphological assessment (Rinaldi et al., 2016, 2017). The identification of the three classes (water, vegetation, and sediments) in time (e.g., monthly, yearly) and space (river reach scale) allow to measure changes and interaction between them and extract hydromorphological indicators useful for river

monitoring. This can be done, for example, by calculating the ratio of "units submerged" to "units submerged + units emerged" and creating a spatio-temporal matrix of these values. Such analysis provides insights on the river dynamics and planimetric variations, useful for understanding past, current, and potential future river processes. In addition, exploiting a time series analysis of derived satellite data, the frequency map of each macro-unit class can be calculated. These products can be used for a variety of purposes, ranging from decision-making support to pre- and post-event assessments (e.g., assessing the impact of natural events such as droughts and/or floods).

Figure 3, boxes A and B, presents a semi-quantitative assessment of the impact of the severe drought that affected North-western Italy between late 2021 and April 2023, focusing on a portion of the Po River. Specifically, it examines the 'water frequency' map calculated as the average frequency of the water component through time. By comparing the period of January-July 2022 vs January-July over the years 2016-2021, the effect of the severe drought is shown by the decrease in the number of pixels belonging to the "water" class, compared with what was observed in the previous years. While not shown here, similar analyses and applications can be conducted for the "vegetation" and "sediment" classes. For example, the map of "exposed sediments" can be used to perform correlation analyses between satellite data and sediment characteristics (e.g., sediment size and shape, sediment provenance, lithology). A first application of this kind can be found in Marchetti et al. (2022), which shows the potential of Sentinel-2 multispectral data to derive gravel vs sand-dominated river bars.

This study demonstrates the potential of Copernicus satellite data, including both multispectral (Sentinel-2) and radar (Sentinel-1), to support the hydromorphological characterization of watercourses as well as their integration with high resolution UAS data.

On the one hand, the EU Copernicus program offers freely available, long-term satellite data with reasonably frequent acquisition, enabling the derivation of statistically robust results for river monitoring. Notably, these Sentinel-1 and Sentinel-2 missions have seen spatial resolution improvements in recent years, now ranging from 10 to 20 meters.

Despite these enhancements, the current resolution still limits the applicability of the methodology to watercourses with active channels wider than 50 meters. Unfortunately, this excludes most mountain river basins. Higher resolution sensors, such as those offered by private satellite companies or CubeSats (with resolution up to 30 cm), exist but come at a significant cost. In mountain river basins, UAS acquisitions remain a valuable tool for detailed hydromorphological characterization, particularly before and after specific events. The future availability of higher satellite resolution imagery, such as those coming from the IRIDE satellite constellation (Table 1) for the Italian public administration, promises to significantly improve the recognition of river forms and processes through satellite-based algorithms.

5.2 Water turbidity monitoring on the Bode River

The Bode catchment is one of the meteorologically and hydrologically best-instrumented meso-scale catchments in Europe providing high resolution observation on water quantity and quality (Zacharias et al., 2011; Wollschläger et al., 2017). Within this catchment, the Meisdorf station has been chosen to assess the

viability of camera systems for continuous water quality monitoring. In addition, targeted experiments have been conducted inducing controlled changes in water turbidity to test potentials of camera systems.

Preliminary results of the long-term study are given in figure 4 which describe the turbidity measurements in terms of Nephelometric Turbidity Units (NTU) obtained by a submerged turbidimeter and the water level gauges. Within this figure, we have selected 7 significant images obtained by the trap camera which provides evidence of the relative changes of the water body colours at different levels of turbidity (Miglino et al., 2022).

In addition, supplementary tracer experiments were conducted adopting kaolin clay (a commonly used turbidity standard, which is readily available, safe, and cost-effective clay mineral) upstream of the monitored river cross-section. Figure 5 illustrates one of these simulated events. During these events, various sensors, including optical cameras, multispectral cameras, and a drone equipped with an optical camera, were employed to capture different perspectives of the synthetic turbidity event. The data collected from these cameras were compared with the records from existing turbidity sensors at the Selke River cross-section.

Figure 5 provides an initial overview of the experiment results, emphasizing the phases of the experiment captured by RGB cameras. However, the UAS imagery offers a richer and more informative depiction of the water dynamic response, highlighting the advantages of zenithal camera positioning for comprehensive observations.

This field campaign proved invaluable for developing and refining the image processing methodology. Integrating field results allowed us to test the camera monitoring system in diverse environmental and hydrological settings. This enabled us to generalize the procedure for potential application to other water quality parameters exhibiting spectral response changes.

These experiments illuminated both the advantages and challenges of camera systems. Though they offer a wealth of information, they can also be susceptible to significant noise and difficulties. While initial findings clearly demonstrate the system's ability to provide insights into river system trends and dynamics, several practical issues require resolution:

- 1) Significant variations in illumination cause notable differences in spectral signatures across days and seasons.
- 2) Changing illumination direction and resulting tree shadows alter the scene during the day.
- 3) Flow conditions and ripples may affect the spectral signature.
- 4) Suspended sediment colour variations can also impact the camera signal.
- 5) Wind, human activity, or animal presence might disrupt camera positioning or even induce vibrations.
- 6) Shallow rivers' backgrounds, particularly the cross-section, can be challenging.

These critical factors should be considered to refine camera systems' effectiveness establishing a robust workflow for high-quality measurements. This might involve optimizing camera angles, implementing effective image filtering and calibration, securing camera stability, and providing comprehensive training for floating object detection.

5.3 Water quality monitoring on the Sarno River

The Sarno River is a formidable challenge: developing pioneering monitoring systems for Europe's most polluted river, situated within a complex and socioeconomically disadvantaged setting. The site has been subject to political disputes over remediation measures for a long time. It is considered the most polluted river in Europe and one of the ten most polluted rivers in the world. The river drains a watershed area of 540 km² densely populated with heavy agricultural and industrial activities. The distribution of different activities in the basin is clustered, which leads to strong spatial variability and temporal fluctuations of environmental conditions (Montuori et al., 2013; Cicchella et al., 2014; Baldantoni et al., 2018).

Figure 6 shows four images of the Sarno River at Scafati in different conditions: ordinary condition, dense presence of macroplastic (Polystyrene foams) and organic material, presence of plastic elements, and presence of foam on the surface.

UAS and camera images have been instrumental in capturing a comprehensive view of water quality dynamics, encompassing factors such as turbidity levels, plastic presence, and pollutant transport. Notably, images offer a distinct advantage in their ability to reveal the spatial distribution of pollution. This capability enables us to precisely identify the sources of specific contaminants and accurately depict the intermingling processes between polluted and clean water. In this context, Figures 7 and 8 below provide a detailed representation of the mixing of contaminated water. Figure 7 illustrates the confluence of the Nocerino tributary into the Sarno River using thermal cameras, while Figure 8 depicts the outlet of the Sarno River using a turbidity index.

Figure 7.A presents an RGB aerial image of the confluence of the Nocerino tributary and the Sarno River. The image clearly shows the distinct chromatic differences between the two water bodies, further confirmed by the thermal image in Figure 7.B. The higher temperature visible in the Nocerino tributary suggests a substantial inflow of sewer water, likely causing the thermal contrast.

Downstream from the confluence, the river's turbidity noticeably increases, serving as a strong indicator of aquatic ecosystem pollution. As the river approaches its mouth at the Tyrrhenian Sea, a clear blending of this concentrated pollutant plume with seawater becomes evident. This mixing process is vividly illustrated in Figure 8.B, where the turbidity index (NDTI) derived from RGB UAS imagery visually depicts this phenomenon.

Besides the characteristics of the fluid, feature detection and classification algorithms can support the identification of floating objects on the water surface. In fact, van Lieshout et al. (2020) compared automatic procedure with manual counting and obtained promising results in several monitored sites in Jakarta (Indonesia). They developed a deep learning algorithm that allowed them to estimate plastic density with a precision of 68.7%, when compared to human detection (see Figure 9).

In the literature, numerous well-established algorithms for feature detection and classification are available. One notable example is the YOLO (You Only Look Once) algorithm, initially developed by Redmon et al. (2016). This algorithm stands out for its exceptional computational efficiency, making it a highly promising tool.

The effectiveness of YOLO in plastic detection has been explored through the application of YOLOv7 (Wang et al., 2023), which demonstrated good performance in identifying plastic objects, as illustrated in Figure 10. Nonetheless, there remains room for improvement in enhancing its performance, especially when it comes to

adapting to the unique environmental characteristics of plastic objects. This can be achieved by incorporating environment-specific plastic training data that encompasses a wide range of colours and geometric variations. In this context, the activities along the Sarno river serve as an ideal source of new training data. Leveraging this data will contribute to optimizing the utilization of a tool that has rapidly evolved into an indispensable and versatile resource for a multitude of applications.

In order to improve the training of these algorithms, there is a need for a large amount of training data which could be provided by any volunteer. With this aim, we are promoting the construction of an imagery repository for YOLO training collecting images of rivers with the presence of different pollutants or floating materials. This activity will be carried out within the project RiverWatch and is already available at the following link:

<https://forms.gle/WHvGvqQc6p3zdEuN9> .

It is worthy to mention that several repositories of river images have been already implemented and these may represent a good starting point in future studies on plastic transport and identification of other pollutants. The most recent repositories developed within different EU or national projects are listed in Table 2.

6 Discussion

This paper addresses the challenges of hydrological monitoring, encompassing issues such as data scarcity, spatial and temporal variability, and the growing demand for water resource monitoring. These challenges are further exacerbated by climate change, which will have significant environmental and socioeconomic impacts. Our analysis emphasizes the need for high-resolution observations in this rapidly evolving environment, where river dynamics and water quality undergo constant changes. Developing new tools capable of providing a holistic picture of the river's state is crucial for securing water supply and ensuring environmental safety.

In this context, the international community is actively exploring the potential of innovative approaches and technologies, including satellite-based technologies, unmanned aerial systems (UAS), camera systems, big data analytics, sensor networks, and citizen science. The integration of multiple data sources and the utilization of innovative approaches offer new prospects for addressing data scarcity, enhancing spatial and temporal resolution, and advancing our comprehension of complex hydrological processes.

Combining remote sensing data with machine learning (ML) holds great potential for creating more comprehensive river monitoring systems (Maier & Keller, 2019; Arias-Rodriguez et al., 2020). Global experiments across diverse basins highlight both challenges and advantages of this approach. We have actively participated in and developed initiatives showcasing the convenience and wide applicability of remote sensing coupled with image processing techniques. These experiences lay the groundwork for a new generation of sensors—from RGB to hyperspectral and thermal—unlocking the power of image processing to extract diverse hydraulic, hydromorphological, and water-quality-related parameters.

Reviewing existing activities and research reveals critical limitations and challenges in using innovative technologies for river monitoring:

1. Developing systems flexible enough to handle diverse hydraulic, environmental, and climatic conditions remains a key challenge. Different river cross-sections within the same basin can have unique characteristics

due to upstream features or local specifics. This requires testing in diverse environments and sharing experiences and data openly with the research community.

2. Although satellites are powerful tools for hydrological and hydromorphological monitoring, their limited spatial resolution often hinders their use in narrower river networks worldwide. Proximity sensors, however, offer a promising alternative to overcome this limitation.

3. Image-based techniques are susceptible to local factors like illumination, surrounding environment, tree shadows, and cross-section background. Developing standardized pre-processing techniques is crucial to ensure consistent data quality and maximize information in each survey or measurement.

4. Individual pollutants and their combinations within a specific location can complicate the interpretation of spectral signatures in any type of imagery, posing a significant challenge that needs to be addressed.

5. The substantial volume of data collected by these sensors may lead to information overload, which may not always be useful for specific purposes. Consequently, there is a need to synthesize the information and develop meaningful metrics capable of retaining essential data.

6. New techniques based on camera use or citizen participation can be susceptible to significant disturbances and procedural errors, potentially degrading the quality of the collected information. Therefore, establishing standardized protocols and sustainable systems for using such methods is critical.

There are still many issues that require further investigation to bring these methods to operational use. However, numerous opportunities exist for advancing hydrological monitoring using these innovative techniques. Here, we highlight some of the key opportunities:

1. Affordable Commercial Devices: The use of image-based techniques could lead to the development of cost-effective commercial devices that can be integrated into sensor networks, even in remote regions around the world.

2. Integration of Crowd-Sourced Data: The integration of crowd-sourced data with image processing appears to be a natural progression for the evolution of new monitoring techniques.

3. Citizen-Friendly Tools: Creating user-friendly tools that could potentially be used on smartphones by the public may expand the overall number of hydrological sensors. This could have significant environmental benefits by fostering a community engaged in river monitoring, raising awareness about the state of aquatic ecosystems, and working towards their protection. Additionally, this approach could address some typical limitations often encountered in traditional techniques, such as thefts, vandalism, reliance on a power supply, and high acquisition and running costs.

4. Multipurpose systems: Leveraging these innovative tools can yield a wealth of information regarding watercourse dynamics that surpasses the capabilities of traditional methods. Such data can be invaluable in bolstering the management of river systems, offering a comprehensive depiction of the river's current condition.

Even with limitations and advantages considered, a critical challenge remains shifting from qualitative to quantitative assessments. This transition stands as arguably the biggest hurdle on our journey forward.

7 Conclusion

Advanced hydrological monitoring techniques and technologies enable the timely identification of hydrological and hydromorphological patterns, anomalies, and critical events like floods, droughts, pollution incidents, and debris flows. This facilitates proactive responses and the implementation of appropriate measures to mitigate risks and minimize extreme event impacts on water resources. In addition, integrating citizen involvement in monitoring may encourage responsible behaviour. This information guides the development of adaptive water management strategies that account for uncertainties and future challenges.

Conventional approaches have historically been limited by technical and financial constraints, hindering their ability to provide the needed detail on river dynamics, water conditions, and other hydrological variables. In contrast, the innovative technologies explored in this manuscript offer a range of advantages, presenting opportunities to complement and integrate existing traditional systems. Furthermore, combining remote and proximal sensing with crowd-sourced data holds immense potential for revolutionizing hydrological monitoring, enabling a comprehensive understanding of river systems. AI algorithms could further enhance the insights gleaned from this vast amount of data.

We foresee exciting potential for initiatives like the IAHS MOXXI working group (<https://iahs.info/Initiatives/Working-Groups/MOXXI/>) to accelerate the development of cutting-edge hydrological monitoring solutions. This international collaboration can act as a catalyst for extensive data sharing and foster ground-breaking joint ventures in innovative solutions. Ultimately, such collective efforts could significantly deepen our understanding of the intricate water cycle dynamics, paving the way for truly adaptive, integrated, inclusive, and sustainable water management.

Acknowledgements: The present research has been carried out within the projects: “OurMED : Sustainable water storage and distribution in the Mediterranean”, which is part of the PRIMA Programme supported by the European Union’s Horizon 2020 Research and Innovation Programme under Grant Agreement No 2222 and the PRIN project 2022MMBA8X - “RiverWatch: a citizen-science approach to river pollution monitoring” (CUP: J53D23002260006) funded by the Italian Ministry of University and Research. F.T. acknowledges support from Progetto ECS 0000024 Rome Technopole, - CUP B83C22002820006, PNRR Mission 4 Component 2 Investment 1.5, funded by the European Union – NextGenerationEU.

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Table 1. Initiatives oriented in promoting the use of new technologies in hydrological monitoring.

Project	Period	Aim of the Initiative/Project	Partners	Web-page
Harmonious COST Action (CA16219)	2016-2022	Establish harmonized monitoring practices for UAS-based observations.	26 countries involved and about 200 scientists	www.costharmonious.eu/
CrowdWater	2016-2020 2020-2024	Citizens data collection as a supplement to existing measurements.	University of Zurich, Switzerland	https://crowdwater.ch/
Habitat Mapping	2017-2021	Create innovative services for habitat mapping, including the hydromorphological characterizations of rivers (IRIS), through integration of Copernicus Sentinel images, UAS acquisitions and in situ and modelled data	ISPRA and Italian Space Agency (ASI), Italy	https://www.isprambiente.gov.it/en/projects/emergency-and-environmental-surveillance/asi-ispra-sentinel-collaborative-thematic-platform-for-habitat-mapping-2017-2021?set_language=en
Plastic Spotter	2019-present	Engage citizens to spot, quantify and share data on plastics floating in the canals of Leiden using CrowdWater App.	European Citizen Science Association 12 partners 9 affiliated entities 15 countries	https://eu-citizen.science/project/125
Plastic Origins		Map plastic pollution in European rivers and share data to the public.	Surfrider Foundation Europe, Private and Public organizations in France	www.plasticorigins.eu/
Plastic Pirates – Go Europe!	2016-2024	Uniform plastic collection in riverbanks and areas near bodies of water using a guideline, and data upload.	DLR Projektträger, 12 EU countries	www.plastic-pirates.eu/
Pescadors	2022-2023	Promotion of the scientific method instruction applied to	BETA Technological Center, University of	https://mon.uvic.cat/pescadors-de-plastic/

de Plàstic		plastic pollution in rivers.	Vic-Central University of Catalonia, 4 private and public institutions and organizations in Spain	
EU H2020 MONOCLE	2018-2022	Collection of water quality and color utilizing citizen science.	Plymouth Marine Laboratory Limited, 12 private and public institutions and organizations in EU	https://www.monocle-h2020.eu/
IRIDE Program, under Italian PNRR	2023-2026	Implementation of an end-to-end system made up of a set of sub-constellations of LEO satellites, the operational infrastructure on the ground, and the services intended for the Italian Public Administration, including tools for hydrology and water resource management.	Industrial teams, under ESA contracts, on behalf of the Italian government	https://www.esa.int/Space_in_Member_States/Italy/IRIDE_La_squadra_e_al_completo

Table 2. Public repositories useful for image-based applications.

Locations	Project	Type of Data	Reference
13 case studies across Europe	Harmonious COST Action (CA16219)	Imagery collected for image velocimetry analysis (along with reference data)	https://doi.org/10.4121/uuidd:014d56f7-06dd-49ad-a48c-2282ab10428e
Five stations installed in Jakarta (Indonesia)	Funded by The Ocean Cleanup	Imagery and code for river plastic detection	https://zenodo.org/record/3817117
Saigon river (Vietnam)		Dataset of about 3,688 UAS images	https://data.4tu.nl/datasets/eca46016-b303-4227-9416-e70101dfd413
Saigon river (Vietnam)	River Plastic Monitoring Project (N. 18211)	Floating material annotations	https://data.4tu.nl/datasets/217004df-49d0-4ed7-9367-ed4f131679bd

Figure 1. Stacked area graph with number of rain gauges installed over the Basilicata Region (Southern Italy) between 1916-2020. Initially, a national agency was in charge of the installation and maintenance of the monitoring network (SIMN-*Servizio Idrografico e Mareografico Nazionale*), while several additional networks have been introduced in more recent years addressing different purposes managed by different agencies.

Figure 2. Three examples of cross-sections: (A) Alcantara River (Italy) during a recent drought observed in the summer of 2021; (B) Sarno River (Italy) with macroplastics and organic matter transport; (C) Flood with wood transport in northern Turkey, Ezine River.

Figure 3. Example of satellite-based classification for the Po River (approximately from Morano at Po and Castelnuovo di Scivia, NW Italy) and the Sesia River over the period January-July 2016-2021 using Copernicus Sentinel-2 data. The map in the bottom left corner displays the location of both rivers using star symbols. Water units are represented by shades of blue, vegetation (both riverbed and riparian) by shades of green, and sediments by shades of brown. Boxes A and B specifically focus on the impact of drought on the Po River's water presence. The map on the left of each box shows the frequency of "water" units from January to July across the years 2016-2021, while the map on the right depicts the same for January to July 2022. This comparison allows for easy visualization of changes in water distribution over time.

Figure 4. Turbidity and water level diagram from February 2022- June 2022 with associated camera shooting taken in different levels of turbidity (Meisdorf station on the Selke river - tributary of the Bode River, Germany).

Figure 5. Turbidity diagram measured in February 2023 with associated camera shooting taken in different conditions from the trap-camera and UAS hovering.

Figure 6. Examples of images taken on the Sarno River at Scafati (Italy) where the presence of suspended material is clearly visible: (A) ordinary condition; (B) dense presence of macroplastic (Polystyrene foams) and organic material; (C) presence of plastic elements; (D) presence of foam on the surface.

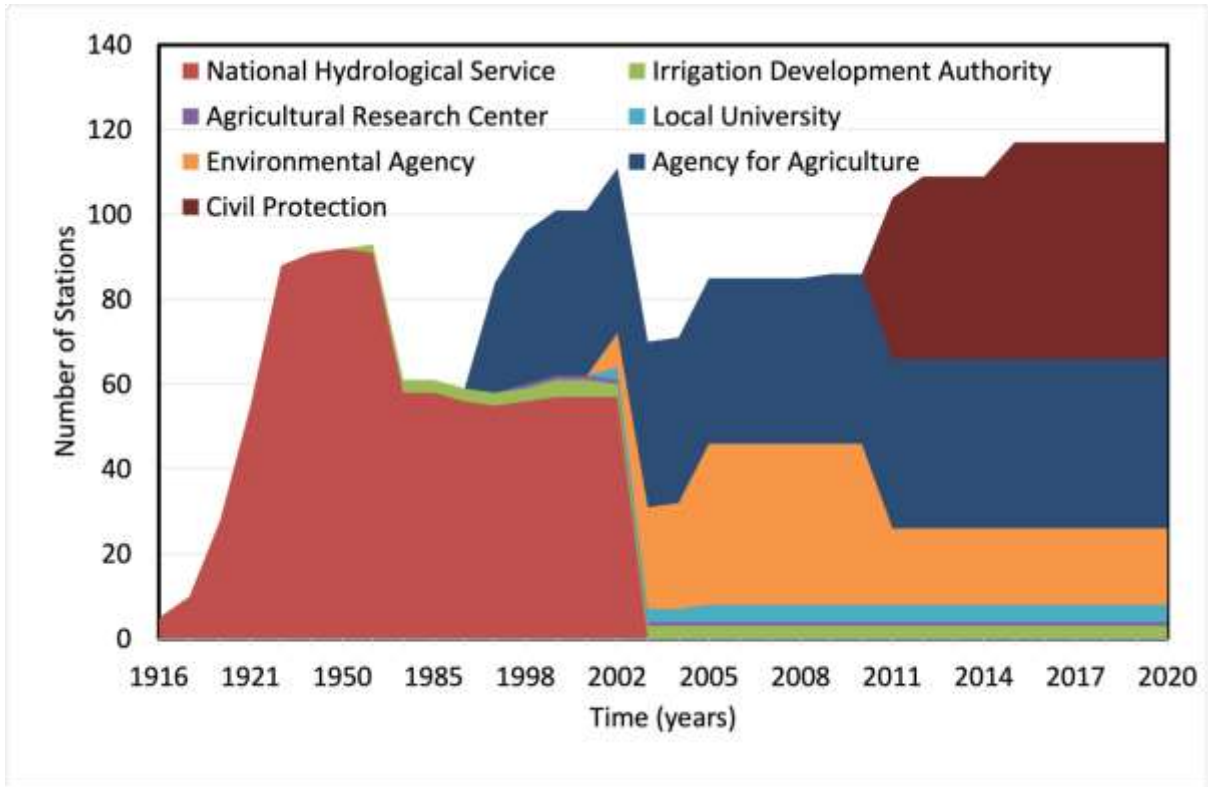
Figure 7. Example of UAS-based imagery obtained with an RGB (A) and a thermal (B) camera at the confluence of the Nocerino River in the Sarno river. Imagery shows a clear flow of pollutants coming from the right tributary.

Figure 8. Outlet of the Sarno river: (A) RGB UAS image, (B) Normalized Difference Turbidity Index, $NDTI=(R-G)/(R+G)$ obtained as the ratio between the red (R) and green (G) bands.

Figure 9. Performances of automatic feature detection for plastic detection applied on five different river locations across Jakarta, Indonesia (taken from van Lieshout et al., 2020).

Figure 10. Sample floating plastic bottles images in the environment with floating macroplastic processed with YOLOv7 algorithm.

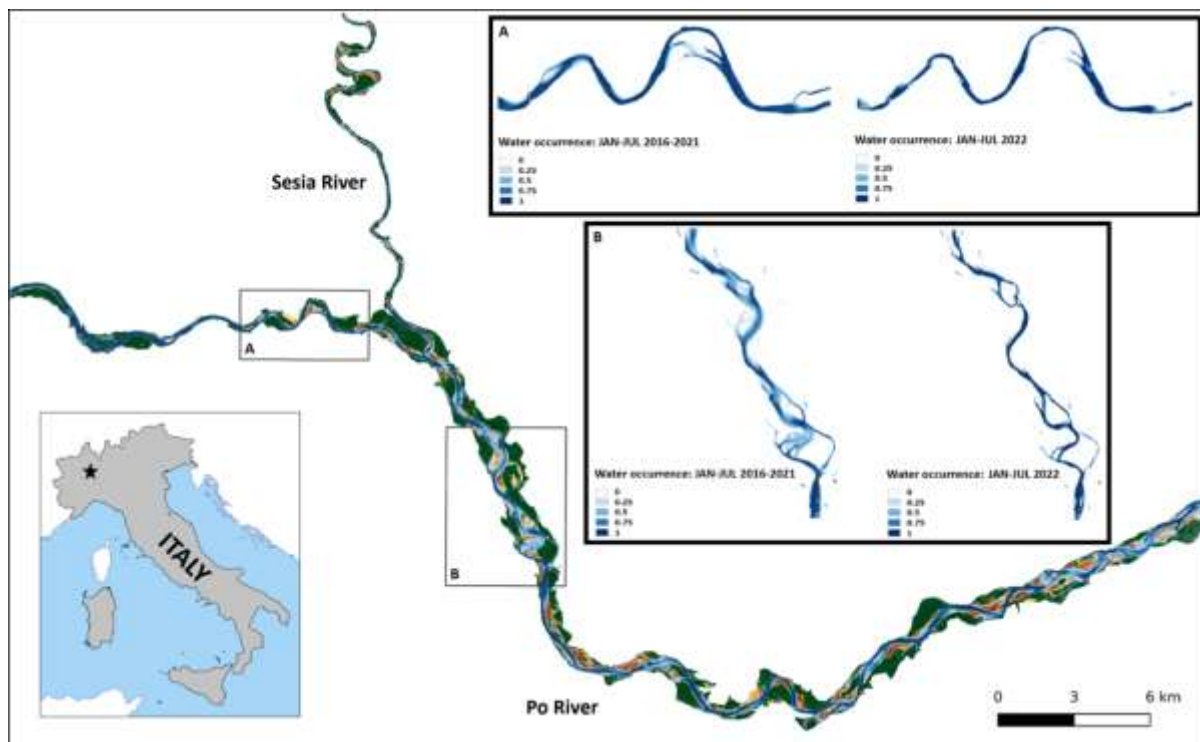
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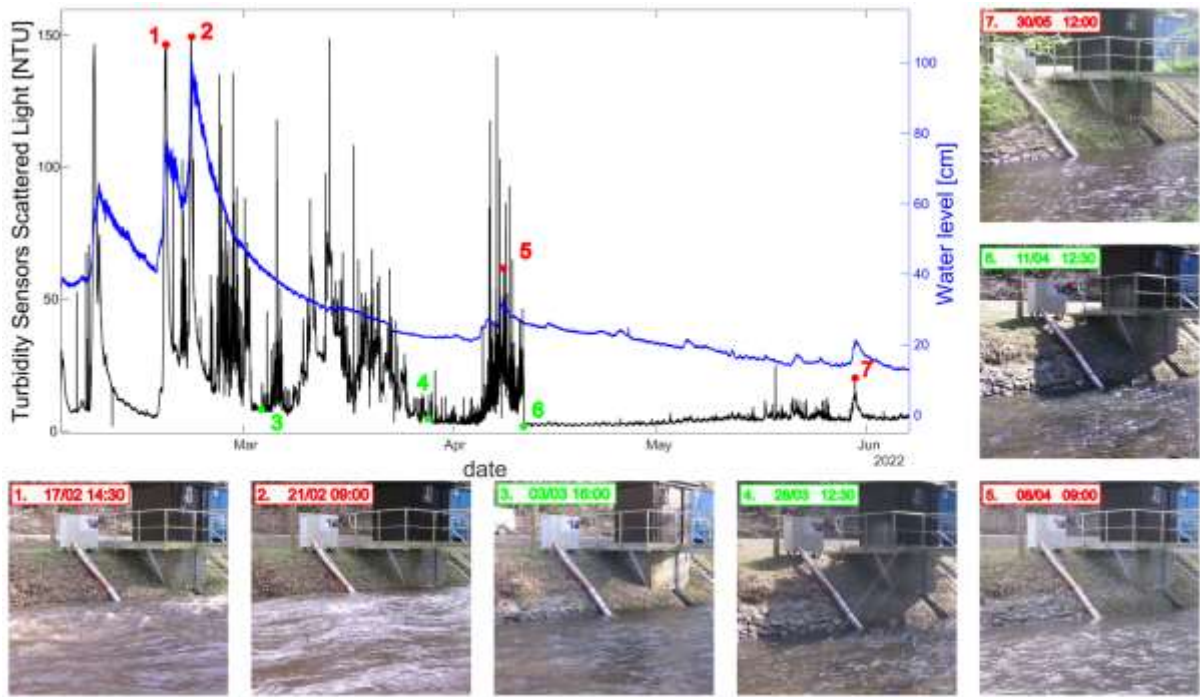
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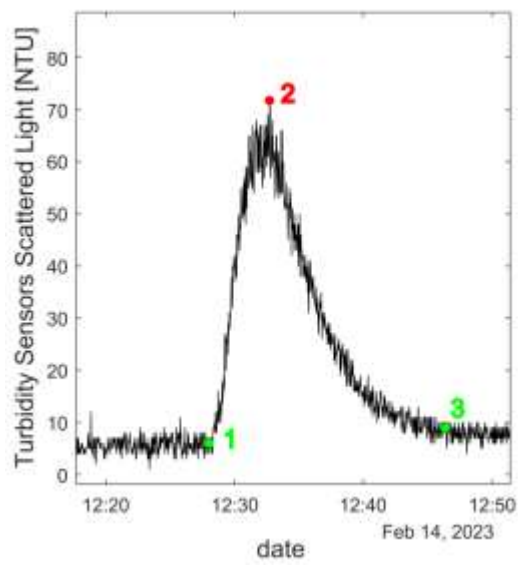
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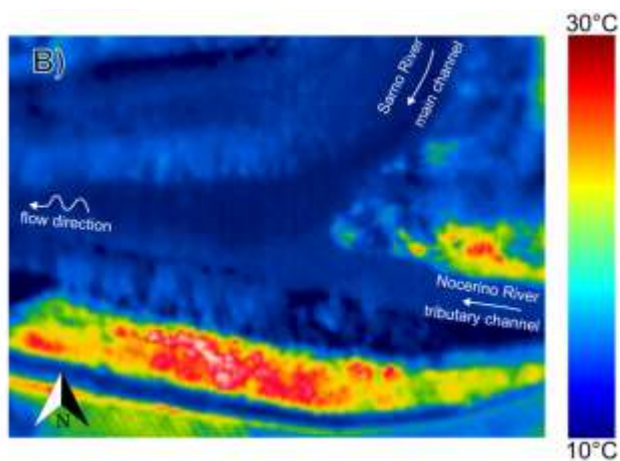
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